

Earthworm's Acetylcholinesterase as Biomarker to Monitor the Effects of Pesticides

Abdel Rahaman, S.M.¹

ABSTRACT

The effect of three insecticides; methomyl (carbamate), dimethoate and malathion (organophosphates) and one copper-containing fungicide (Copper hydroxide) was evaluated *in vivo* on acetylcholinesterase (AChE) isolated from earthworm *Lumbricus terrestris* to achieve a better understanding of AChE responses to agrochemicals in *L. terrestris*. Three values of median lethal concentrations of the four compounds were determined from the acute toxicity (0.1, 0.5 of LC₅₀ and the LC₅₀). The three insecticides at the three levels of LC₅₀ inhibited AChE activity and the inhibition was dose dependent. Methomyl at LC₅₀ level had the highest AChE inhibition (98.9%) after 7 days of exposure to contaminated soil. Copper hydroxide which contains copper element showed induction of AChE activity at the three levels of LC₅₀s. The maximal induction was at 0.5 of LC₅₀ (27.7% followed with 0.1 (20.3%) and the value of LC₅₀ (6.8%) compared with control.

Our results further support the use of AChE as an indicator of pesticide contamination, to be included in a battery of biomarkers for monitoring soil toxicity.

INTRODUCTION

Earthworm constitute 80% of the soil invertebrate biomass in most terrestrial ecosystems of the world (Lee, 1985). They are relatively large, immobile and most easily quantified components of soil biota. They preserve and contribute to the overall productivity of soil ecosystem by maintaining soil structure and regulating the turn over of organic matter through their feeding, casting and burrowing (Dash, 1978; Lee, 1985; Parmelee *et al.*, 1990). Earthworms also form one of the principal source of animal proteins for many predators and occupy a major compartment in the chemical element cycles (Ferriere *et al.*, 1981). Due to their relatively large size, limited rapidity in soil displacement, slow recolonization and beneficial role in agro ecosystems, earthworms are used as an indicator species for monitoring the impact pollutants, changing in soil structure and agricultural practices (Haque and Ebing, 1983; Heimbach, 1985; Panda and Sahu, 1999; Paoletti, 1999; Ping *et al.*, 1999).

Hundreds of manufactured pesticides of different chemical composition are currently used through out world to protect crops against pests. Small amount of applied pesticides reach the target and the rest affect the

non-target organisms (Pimental and Levitan, 1986). For instance, organophosphate and carbamate compounds are generally short-lived in the environment and once ingested or otherwise acquired by an organism, they are rapidly metabolized or excreted (Panda and Sahu, 1997). Organophosphate and carbamate insecticides as neurotoxic agents are known to cause acute toxic effect in earthworm (Scott-Fordsmann and Weeks, 2000; Rao and Kavitha, 2004). The site of action of these insecticides is acetylcholinesterase enzyme (AChE) which hydrolyze acetylcholine in the invertebrate nervous system (Corbett, 1974).

Moreover, copper – containing sprays have been used to control fungal diseases in fruits and vegetable crops (Merry *et al.*, 1983). Copper is the essential element and required by all organisms. However, elevated concentrations of copper are toxic and when found in soil it may lead to a range of effects including reduced biological activity and subsequent loss of fertility (Dumestre *et al.*, 1999).

Poisoning of the nervous system is perhaps the quickest and most effective method of chemically upsetting regular body function (Hoar, 1991). That is why AChE activity is used as a reliable parameter for assessing the poisoning due to pesticides and heavy metals (Reddy and Venugopal, 1993; Sharma *et al.*, 1993; Devi and Fingerman, 1995; Dembele *et al.*, 1999). Few studies were done on the effect of pesticides on kinetic properties of earthworm's AChE, therefore, the objective of the current study was to evaluate the toxicity of some pesticides on earthworm *Lumbricus terrestris in vivo*. The effect of lethal and sublethal concentrations on the activity of earthworm's AChE was also studied.

MATERIALS AND METHODS

Animals

The earthworm (*Lumbricus terrestris*) was collected from the Agriculture Research Center garden, Sabahia, Alexandria. They were carefully brought to the laboratory along with the moist soil within an hour. The worms were acclimatized at the lab conditions (at room temperature and 12 hr light/12 hr dark) in the artificial soil (using an evenly blended dry weight mixture of 20% kaolin clay soil, 70% silica sand, 10% sphagnum peat and 0.3% calcium carbonate, according to the

¹Central Pesticides Laboratory, Sabahia Station Alexandria, Egypt
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Organization for Economic Cooperation and Development; OECD guideline 207 (OECD 1984) before testing.

Pesticides

a) Organophosphate insecticides:

1) Malathion (Camethion® 57% EC)

It was obtained from the Egyptian Center for Agriculture Services.

2) Dimethoate (Camethoate® 40% EC).

It was obtained from Trading Company for Agriculture Requirements.

b) Carbamate insecticide

Methomy (Methocam® 90% SP)

It was obtained from Trading Company for Agriculture Requirements

c) Copper-containing fungicide:

Copper hydroxide (Dakocide® 83.4% WP)

It was obtained from Delta Company for Agriculture Chemicals.

Reagents

All reagents used in the present study were of analytical grade. Acetylthiocholine iodide (ATChI) and 5,5 dithiobis (2-nitrobenzoic acid (DTNB) were purchased from Sigma- Aldrich Chemical Company.

Determination of median lethal concentration (LC₅₀):-

Toxicity experiments were conducted by artificial soil test for 7-days exposure (OECD, 1984). Different concentrations of the tested pesticides were homogeneously mixed with artificial soil. Control was also run in parallel using water alone. Each concentration was replicated three times and placed in plastic containers covered with perforated cloth. Each replicate contain 10 earthworms of approximately equal length (9.52 ± 0.25 cm) and weight (0.27 ± 0.039 g). Mortality percentages were recorded after 7 days of exposure and LC₅₀ values were calculated according to Finney (1971).

Enzyme source preparation:

One gram of anterior adult worm (from segment 1-7) was homogenized in 5 ml of 0.1 M phosphate buffer (pH 7.4) using polytron for 1 min. The homogenate was centrifuged at 5000 rpm for 30 min. at 4 °C. using Beckman GS-6R centrifuge rotor (GA-10). The supernatant was used as an enzyme source.

Protein assay:

The protein content was carried out according to the method described by Lowery *et al.*, (1951), using bovine serum albumin (BSA) as standard.

Cholinesterase activity assay:

The spectrophotometric method of Ellman *et al.*, (1961) was used, with acetyl thiocholine iodide (ATChI) as a substrate. In a typical assay 50 µl of the enzyme source was added to 3 ml of a 1:1 mixture of 2 mM substrate solution and 2 mM dithiodinitrobenzoic (DTNB). The final concentration of substrate and DTNB in the assay mixture was 1 mM. The changes in absorption at 412 nm was monitored on spectrophotometer Spectronic 601. An assay mixture without enzyme was used as the blank.

Kinetic Studies:

Supernatant derived from unexposed earthworms was used to evaluate the maximum velocity of the substrate hydrolysis (V_{max}). Michaelis Menten constant (K_m) was estimated by the double-reciprocal method of Lineweaver and Burk (1934). Optimum protein concentration and optimum incubation time for the earthworm AChE were also determined.

In vivo Effect of Earthworm AChE Activity by Malathion, Dimethoate, Methomyl and Copper Hydroxide.

Earthworms were exposed to the concentrations of 0.1, 0.5 and 1 of the LC₅₀ of malathion, (7.23, 36.14 and 72.27 mg/kg) dimethoate, (1.08, 5.40 and 10.79 mg/kg) methomyl (0.088, 0.44 and 0.88 mg/kg) and copper hydroxide (34.64, 173.21 and 346.42 mg/kg). The enzyme sources and the enzyme assay were done as mentioned before. The enzyme activity in the treatments was calculated as a percent of the enzyme activity in the control.

Results were subjected to analysis of variance (ANOVA) (CoStat statistical Software, 1990). The standard error (SE) of three replications was calculated.

RESULTS AND DISCUSSIONS

In vivo experiment

Toxicity of methomyl, dimethoate, malathion and copper hydroxide on the *L. terrestris* after 7 days of exposure was reported. Median lethal concentrations (LC₅₀s) of the effect of previous pesticides, their confidence limits and slopes were estimated (Table 1). At the LC₅₀ levels, methomyl was the most toxic pesticide (LC₅₀ = 0.88mg/kg of soil), while copper hydroxide was the least toxic one (LC₅₀ = 346.4mg/kg). In other words the three insecticides were more toxic to *L. terrestris* compared with the tested fungicide (copper hydroxide).

Assay of AChE

Three different concentrations of each tested pesticides (0.1, 0.5 of LC₅₀ and the LC₅₀) were estimated from Table 1. The enzyme protein concentration per

Table 1. Toxicity of tested pesticides on the earthworms (*L. terrestris*) after 7 days of artificial soil exposure

Pesticides	LC ₅₀ (mg/kg)	LC ₅₀ (confidence limits)		Slope	Probability
		Lower	Upper		
Methomyl	0.88	0.73	1.06	1.59	0.06
Dimethoate	10.79	8.82	13.18	2.19	0.56
Malathion	72.27	70.07	80.86	4.41	0.97
Copper Hydroxide	346.42	332.90	360.50	5.95	0.84

assay was 125ug. The activity of AChE as change in absorption at 412 nm was determined. The substrate concentration was 2 mM. The kinetics of AChE was also determined. The dissociation constant of the enzyme-substrate complex, defined as k_m (Michaelis constant) was graphically determined by using Lineweaver Burk Plots of reciprocal substrate concentration (1/S) against reciprocal velocity (1/V). The values of k_m and V_{max} were 0.0165 mM and 0.01 m mole / min respectively.

Comparative analysis of the k_m appears lower value than that reported by Caselli *et al.*, (2006) (0.14 mM). This explanation of these results might be due to the highly polymorphic AChE enzymes in most species and the number of genes coding for different isoforms varies between species (Bebiano *et al.*, 2004).

The effect of the four tested pesticides at the three levels of concentrations on earthworms AChE activity after exposure to contaminated soils for 7 days was illustrated (Fig. 1 and Table 2 A,B). Data showed that inhibition of AChE was increased with increasing the concentrations of pesticides. The maximal inhibition of AChE activity was at the LC₅₀ values of insecticides compared with the other two concentrations (0.1 and 0.5 of LC₅₀). Methomyl treatment showed the highest reduction of AChE activity (98.9%) followed by

malathion (93.2%) and dimethoate (89.8%) at the LC₅₀ levels. In contrast, copper hydroxide which contains copper element showed induction of AChE activity at the three levels of LC₅₀, the maximal induction was at 0.5 LC₅₀ (27.7) followed with 0.1 (20.3) and the value of LC₅₀ (6.8) compared with control. These results may refer that still the concentration of copper hydroxide at LC₅₀ value of copper hydroxide activate earthworm's AChE, because many elements play as coenzymes in the biological systems.

The data also reflected that AChE activity was strongly decreased by increasing the concentrations of the tested insecticides, which caused significant inhibition. These data confirm that AChE is the target for methomyl, dimethoate and malathion. The inhibition of AChE activity has been considered as sensitive biomarker to assess pesticide effects on various non-target organisms (Damiens *et al.*, 2004 and Ferrari *et al.*, 2004). The results in line with that reported by Caselli *et al.*, (2006) who found that carbaryl (carbamate insecticide) was able to reduce the earthworm's AChE activity by about 95% when it was used at high concentration (10⁻⁵ M). Moreover, Rao *et al.*, (2003) reported that chlorpyrifos inhibited earthworm's AChE and the inhibition of AChE was a dose and time dependent.

Table (2. A). *In vivo* inhibition of earthworm AChE by methomyl, dimethoate and malathion at different concentrations

Insecticide	Conc.(mg/kg)	S.A ± S.E (ΔO.D/mg protein.min)	Activity (% Control)	% Inhibition
Control	0.0	0.0177 ± 0.0000	100	0.0
Methomyl	0.088	0.0167 ± 0.0003	94.4	5.6 c
	0.44	0.0049 ± 0.0001	28.2	71.8 b
	0.88	0.0002 ± 0.0000	1.1	98.9 a
Dimethoate	1.08	0.0057 ± 0.0003	32.2	67.8 c
	5.40	0.0027 ± 0.0001	15.3	84.7 b
	10.79	0.0018 ± 0.0000	10.2	89.8 a
Malathion	7.23	0.0053 ± 0.0001	29.9	70.1 c
	36.14	0.0037 ± 0.0001	20.9	79.1 b
	72.27	0.0012 ± 0.0000	6.8	93.2 a

Numbers within the same insecticide followed by the same letter are not significantly different.

Table (2. B). *In vivo* induction of earthworm AChE by cupper hydroxide at different concentrations

Conc.(mg/kg)	S.A ± S.E (Δ O.D/mg protein.min)	Activity (% Control)	% Induction
0.0 (control)	0.0177 ± 0.0000	100	0.0
34.64	0.0213 ± 0.0002	120.3	20.3 b
173.21	0.0266 ± 0.0000	127.7	27.7 a
346.42	0.0189 ± 0.0001	106.8	6.8 c

Numbers followed by the same letter are not significantly different.

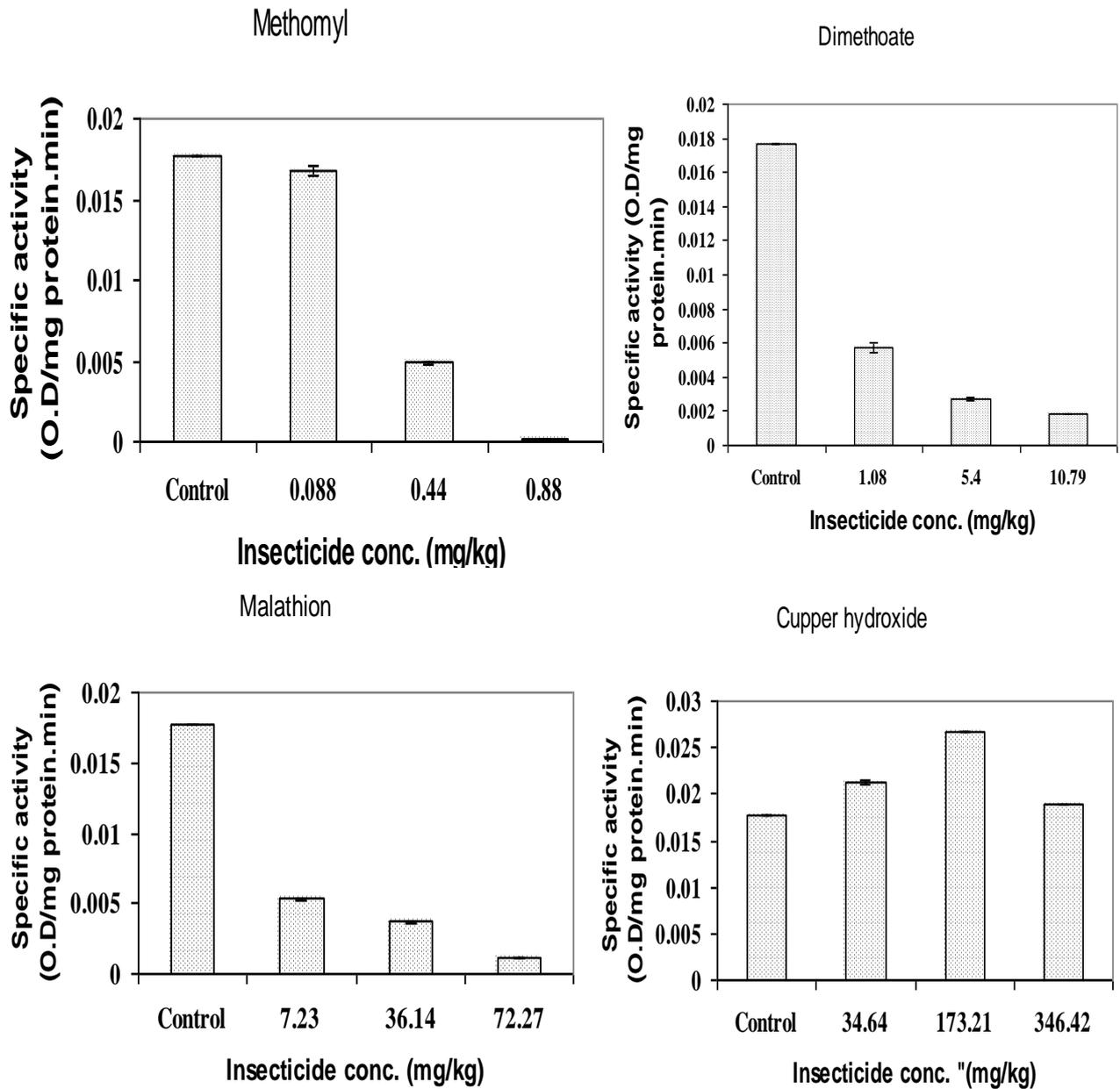


Fig 2. *In vivo* effect of certain pesticides on the AChE activity of earthworm

Biomarker responses have great potential for assessing site pollution because they integrate a wide array of environmental toxicological and ecological factors that control and modulate exposure to contaminants, however many non-pollution related variables may interfere with biomarker response (Arnaud *et al.*, 2000). The need to increase our knowledge of biomarker responses in earthworms has been stressed (Kammenga *et al.*, 2000, Sanchez-Hernandez, 2006, Gambi *et al.*, 2007 and Rault *et al.*, 2007).

In conclusion, the activity of earthworm's AChE that assessed *in vivo* was dose dependently inhibited by carbamate or organophosphate compounds which might acts as competitive inhibitors. AChE activity isolated from earthworms could be used as a biomarker for insecticide contaminants. Our results further support the use of AChE as a biomarker of pesticide contamination, and could be used for monitoring soil contamination.

REFERENCES

- Arnaud, C.; Saint-Denis, M.; Narbonne, F. and Soler, R. D., 2000. Influences of different standardized test methods on biochemical responses in the earthworm *Eisenia foetida andrem*. *Soil Biol. Biochem.* 32, 67-73.
- Bebiano, M.J.; Geret, F.; Hoarau, P.; Serafim, M.A.; Coelho, M.R., Gnassia-Barelli, M. and Romeo, M., 2004. biomarkers in *Ruditapes decussates*: a potential bioindicator species. *Biomarkers* 9, 305-330.
- Caselli, F.; Gastaldi, L.; Gambi, N. and Fabbri, F., 2006. *in vitro* characterization of cholinesterases in the earthworm *Eisenia andrei*. *Comp. Biochem. Physiol. C* 143, 416-421.
- Corbett, J.R., 1974. *The Biochemical Mode of Action of Pesticides*. Academic Press, London.
- CoStat Statistical Software, 1990. microcomputer program analysis version 4.20, CoHort Software, Berkeley, CA.
- Damiens, G.; His, F.; Gnassia-Barelli, M.; Quiniou, F. and Romeo, M., 2004. Evaluation of biomarkers in oyster larvae in natural and polluted conditions. *Camp. Biochem. Physiol., C* 139, 121-128.
- Dash, M.C., 1978. Role of earthworms in decomposer system. In: Singh, J.S., Gopal, B. (Eds.), *Glimpses of Ecology*. International Scientific Publications, New Delhi, pp. 399-406.
- Dembele, K.; Haubruge, E. and Gasper, C.H., 1999. Recovery of acetylcholine esterase activity in the common carp (*Cyprinus carpio* L.) after inhibition by organophosphate and carbamate compounds. *Bull. Environ. Contam. Toxicol.* 62, 731-742.
- Devi, M. and Fingerman, M., 1995. Inhibition of cholinesterase activity in the central nervous system of red swamp cary fish, *Procambarus clarkii* by mercury, cadmium and lead. *Bull. Environ. Contam. Toxicol.* 55, 746-750.
- Dumestre, A.; Sauve, S.; McBride, M., Baveye P. and Berthelin, J. 1999. Copper speciation and microbial activity in long-term contaminated soils. *Arch. Environ. Con. Tox.*, 36, 124-131.
- Ellman, G.L.; Courtney, K.D.; Andres, V. and Featherstone, R.M., 1961. A rapid calorimetric determination of acetylcholinesterase activity. *Biochem. Pharmacol.*, 7, 88 - 95.
- Ferrari, A.; Anguiano, O.L.; Solcone, J., Venturino, A. and Pechen de D'Angelo, A.M., 2004. Different susceptibility of two aquatic vertebrates (*Oncor-* and *Bufo arenarum*) to azinphosmethyl and carbaryl. *Comp. Biochem. Physiol., C* 139, 239-243.
- Ferriere, G.; Fayolle, L. and Bouche, M.B., 1981. Un nouvel outil, essentiel pour l'ecophysilogie et l'ecotoxicologie, l'elevage des lombriciens en sol artificiel. *Pedobiologia* 22,96-201.
- Finney, D.J., 1971. *Probit Analysis*. Cambridge University Press, Cambridge.
- Gambi, N.; Pasteris A. and Fabbri E. 2007. Acetylcholinesterase activity in the earthworm *Eisenia andrei*. *Comp. Biochem. Physiol. C* in press.
- Haque, A. and Ebing, W., 1983. Toxicity determination of pesticides to earthworms in the soil substrate. *Z. Pflanzenkrank Pflanzenschutz* 90, 395-408.
- Heimbach, F., 1985. Comparison of laboratory methods using *Eisemia fetida* and *Lumbricus terrestris* for the assessment of hazard of chemicals to earthworms. *Z. Pflanzenkrank Pflanzenschutz* 92, 186-193.
- Hoar, W.S., 1991. *General and Comparative Physiology*. Prentice Hall of India Pvt. Ltd., New Delhi, India.
- Kammenga, J.E.; Dallinger, R.; Donker, M.R.; Kohler, B.R., Simonsen, V., Trieskorn, R. and Weeks, J.M., 2000. Biomarkers in terrestrial invertebrates for ecotoxicological soil risk assessment. *Rev. Environ . Contam. Toxicol.* 164, 93-147.
- Lee, K.E., 1985. *Earthworms, their Ecology and Relationships with Soil and Land Use*. Academic Press, Sydney.
- Lineweaver, H. and Burk, D., 1934. Determination of enzyme dissociation constants. *J. Ann. Chem. Soc.* 56, 658-666.
- Lowery, O.H.; Resebrough, N.J.; Farr, A.L. and Raisdall, R.J., 1951. Protein measurement with Folio phenol reagent. *J. Boil. Chem.* 193, 265-275.
- Merry R.H.; Tiller, K.G and Alston A.M. 1983. Accumulation of copper, lead and arsenic in some Australian orchard soils. *Aust. J. Soil Res.* 21, 549-561.
- OECD, 1984. *Earthworms, acute toxicity testes*. In: Organization for Economic Cooperation and Development (Ed.) *Guidelines for Testing of Chemicals*. No. 222. Paris, France.
- Panda, S. and Sahu, S.K., 1997. Recovery of respiratory and excretory activity of *Drawida willsi* (Oligochaeta) following application of malathion in soil. *J. Ecobiol.* 9 (2), 97-102.

- Panda, S. and Sahu, S.K., 1999. Effects of malathion on the growth and reproduction of *Drawida willsi* (Oligochaeta) under laboratory conditions. *Soil Biol. Biochem.* 31, 363-366.
- Paoletti, M.G., 1999. The role of earthworms for assessment of sustainability and as bioindicators. *Agric. Ecosyst. Environ.* 74 (1-3), 137-156.
- Parmelee, R.W.; Beare, M.H., Cheng, W.; Hendrix, P.F., Rider, S.J. and Coleman, D.C., 1990. Earthworms and enchytraeids in conventional and no-tillage agroecosystems: A biocide approach to assess their role in organic matter breakdown. *Biol. Fertile. Soils* 10, 1-10.
- Pimentel, D. and Levitan, L., 1986. Pesticides; amount applied and amount reaching pests. *Bioscience* 36, 86-91.
- Rao, J.V.; , Pavan Y. Surya., Madhavendns, S.S., 2003. Toxic effect of chlorpyrifos on morphology and acetylcholinesterase activity of the earthworms *Eidensni foetida*. *Ecotoxicol. Environ. Saf.* 54, 206 301.
- Rao, J., Venkateswara and P. Kavitha 2004. Toxicity of azodrin on the morphology and acetylcholinesterase activity in the earthworm, *Eisenia foetida*. *Environmental Research*, 96, 323-327.
- Rao, J., Venkateswara, Pavan Y. Surya and S.S. Madhavendrab, 2006. toxic effect of chlorpyrifos on morphology and acetylcholinesterase activity in the earthworm, *Eisenia foetida*. *Eco. Toxicology and Environmental Safety* 54 (296-301).
- Rault, M.; Mazzia C. and Capowiez Y. 2007. Tissue distribution and characterization of cholinesterase activity in six earthworm species. *Comparative biochemistry and physiology part B* 147, 340-346.
- Rault, M.; Mazzia C. and Capowiez Y. 2007. Tissue distribution and characterization of cholinesterase activity in six earthworm species. *Comparative biochemistry and physiology part B* 147, 340-346.
- Sanchez-Hernandez J.C. 2006. Earthworm biomarkers in ecological risk assessment. *Rev. Environ. contam. Toxicol* 188, 85-126 Links.
- Scott- Fordsmand, J.J. and J.M. Weeks, 2000. Biomarkers in earthworms. *Rev. Environ. Contam. Toxicol.* 165, 117-159.
- Sharma, B., Gopal, K. and Y.P., Khanna, 1993. Interaction of carbaryl with acetylcholinestrase of teleost *Clarias batrachus*. *Toxicol. Environ. Chem.*39, 147-152.

الملخص العربي

الأسيتايل كولين إستيريز لديدان الأرض كمؤشر حيوى لرصد تأثيرات المبيدات

صفاء مصطفى عبد الرحمن

التشبيط مع زيادة التركيز المستعمل فى الثلاث مبيدات الاولى السابقة الذكر. وأظهرت النتائج أيضا أن أعلى تشبيط للإنزيم كان بإستخدام الجرعة القاتلة لـ 50% من الديدان وذلك بعد 7 أيام من تعرض دودة الارض للتربة الملوثة بمبيد الميثوميل (98.9%).

أما المبيد الفطرى كبر هيدروكسيد (داكوسيد) والذى يحتوى على عنصر النحاس فقد أظهر تنشيط لإنزيم الأسيتايل كولين إستيريز عند الثلاث مستويات المختبرة من قيم LC_{50} وكان أعلى تنشيط عند إستخدام نصف جرعة LC_{50} متبوعا بقيمة عشر LC_{50} وأخيرا قيمة LC_{50} (27.7%، 20.3% و 6.8) على التوالى وذلك بالمقارنة بالكنترول.

هذه النتائج تعضد إستخدام إنزيم الأسيتايل كولين إستيريز كمؤشر حيوى لرصد تلوث التربة بالمبيدات ضمن المؤشرات الحيوية الاخرى مع إعتبار ديدان الأرض كائن دليلى.

تم تقدير التأثير الداخلى لثلاث مبيدات هى الميثوميل (من مجموعة الكاربامات) والداى ميثويت والملاثيون (من مجموعة الفوسفور العضوية) ومبيد من المبيدات الفطرية المحتوية على عنصر النحاس (كبر هيدروكسيد) على نشاط إنزيم الأسيتايل كولين إستيريز المعزول من دودة الارض لمعرفة أفضل إستجابة لإنزيم الأسيتايل كولين إستيريز للكيمواويات الحقلية فى ديدان الارض.

وقد أستخدم ثلاث تركيزات من الجرعات القاتلة لـ 50% من الديدان (LC_{50}) للاربعة مركبات والى تم تقديرها من إختبارات التقييم الحيوى لدراسة تأثيرها على نشاط الإنزيم وهى قيمة LC_{50} ، ونصف وعشر هذه الجرعة.

وقد أدت هذه المبيدات بالثلاث تركيزات المستخدمة و المختبرة الى تشبيط نشاط إنزيم الأسيتايل كولين إستيريز وقد ارتبطت شدة